

Ecological survey of Donald Creek and Otauira Stream, 2016

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Cover photo: Fyke net in Otauira Stream.

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Contents

1 Introduction

1.1 Background

Featherston Wastewater Treatment Plant (WWTP) discharges treated wastewater to Donald Creek. Donald Creek enters the Otauira Stream about 2.5km downstream of the discharge, and Otauira Stream flows another 2.6 km before entering the northern end of Lake Wairarapa. The South Wairarapa District Council (SWDC) is undertaking a programme to upgrade the WWTP to partial land disposal but there is likely to still be a discharge to Donald Creek for a period of time and under wet conditions that typically occur during winter and spring.

River Lake Ltd was commissioned to undertake an ecological survey of Donald Creek and Otauira Stream to assess the effects of the discharge during spring conditions and to inform the assessment of environmental effects. This report discusses the results of ecological monitoring undertaken in early October and early November 2016.

1.2 Past ecological surveys of the stream

Ecological surveys of Donald Creek were undertaken at two sites in April 2010 and five sites in March 2013 to assess the effects of the WWTP (Coffey 2010, Coffey 2013). Shading by trees was thought to mask the full effect of the nutrient enrichment so additional sites were sampled in 2013. Also, some willows were killed during 2013 to make the stream more open. The key findings of these surveys were:

- The WWTP discharge caused conspicuous change in colour at the point of discharge.
- Some heterotrophic growths were found downstream (ca.5% cover for both surveys).
- Periphyton cover and biomass was significantly higher downstream of the discharge (about 15 % cover of long filamentous algae in 2010). In 2013 the difference in periphyton cover and biomass was particularly apparent at the unshaded sites and very little periphyton was present at the shaded sites.
- In 2010, the periphyton community was dominated by *Fragilaria* sp., *Gomphonema* sp. and *Phormidium* sp. upstream of the discharge, while *Stigeoclonium* sp. dominated downstream. In 2013 the periphyton community was dominated by *Stigeoclonium* sp. upstream and *Fragilaria* sp. downstream.
- All macroinvertebrate metrics indicated a statistically significant decline in water quality downstream of the discharge.

Planktonic green algae were common in the periphyton samples and Daphnia was common in invertebrate samples – both indicative of a WWTP pond discharge. In 2013 sites immediately downstream of the discharge had planktonic algae from the oxidation ponds as a scum on substrate and moss.

Forbes (2013) undertook ecological monitoring of Donald Creek about monthly between November 2012 and April 2013. This consisted of water quality samples and visual assessments of periphyton cover and analysis of periphyton relative abundance. Sites were located about just upstream, about 60m downstream and about 160m downstream. Key findings of the monitoring were:

A reduction in substrate size downstream with more sand and silt at the downstream sites.

- All sites were clear of periphyton during November. During summer the amount of periphyton cover increased and this was particularly noticeable at the downstream sites. During April, 'sludge' (medium mats) covered about 10% of the bed and coarse filamentous green algae covered almost 20% of the bed at the 160m downstream site.
- Periphyton biomass at the 160m downstream site during February was 138 mg chlorophyll a/m^2 and 42 g AFDM/m²; during April it was 91 mg chlorophyll- a/m^2 and 17.5 g AFDM/m². Upstream samples were not collected.
- The downstream sites had substantially lower water clarity (about 0.6m compared to 2.6m upstream), higher suspended solids (ca.6 to 9 mg/L compared to 4 mg/L upstream), higher BOD (about 4 to5 mg/L compared to <1 mg/L upstream), noticeably higher soluble inorganic nitrogen (SIN) (1 mg/L compared to 0.6 mg/L upstream) and dramatically higher dissolved reactive phosphorus (ca. 0.3 mg/L compared to ca. 0.01 mg/L upstream).

In previous reports Otauira Stream is also known as Abbot Creek; Donald Creek is also known as Boar Creek.

2 Methods

2.1 Timing and flow conditions

Ecological surveys were undertaken in Donald Creek and Otauira Stream twice during spring. The surveys occurred on 10th -11th October 2016 and on 1 November 2016; and the stream flows at the time of the surveys were 320 L/s and 259 L/s respectively. A large flood had occurred on 18 September (3450 L/s) – about four weeks prior to the October survey (Figure 2.1). Debris from the flood was visible in the stream at the time of the survey.

The water level in the stream at the time of the survey was typical of spring base flow conditions (the median flow in October for an 11 year synthetic flow record (2005-2016) was 332 L/s). The annual median flow in Donald Creek upstream of the Featherston WWTP discharge is 241 litres per second (Butcher 2016).

Figure 2.1: Daily average flow in Donald Creek and time of the stream surveys 10 -11 October and 1 November 2016.

2.2 Sites

The ecological survey sampled four sites on Donald Creek and two sites on Otauira Stream. The description and location of the sites are show in Figure 2.2 and Table 2.1). The sites on Donald Creek are consistent with those sampled in 2013 by Coffey (2013). The sites 25m upstream of the discharge and 60m downstream of the discharge were within a remnant of protected bush, the sites 100m upstream and 650m downstream were surrounded by pasture. The 650m downstream site had noticeably slower flows and smaller substrate than the other sites. A small tributary (Longburn water race) enters Donald Creek from the true left about 430m downstream of the discharge.

The sites on Otauira Stream were located upstream and downstream of the confluence with Donald Creek. These sites had not been previously sampled by Coffey (2013).

The fish survey on Donald Creek placed nets over a 140 m reach from about 90m upstream to 60m downstream of the discharge. The fish survey on Abbot Creek placed nets over 200 m reach from about 100 m upstream to 100 m downstream of the confluence.

Table 2.1: Location of sample sites for ecological monitoring

Figure 2.1: Location of sample sites on Donald Creek and Otauira Stream (also known as Abbot Creek). The bottom map shows a detail of Donald Creek.

2.3 Habitat

In-stream habitat quality was recorded to help interpret data from macroinvertebrate and periphyton sampling. At each site, habitat was assessed using the National Rapid Habitat Assessment Protocol (Clapcott 2015) and a Habitat Quality Score (HQS) calculated. In each sample reach observations were made of water depth, stream bed particle size, discolouration of water column, and evidence of foaming.

Water depth and velocity was measured mid-stream of run sections using the ruler method described in Harding et al. (2009).

The particle size distribution of substrate on the stream bed was assessed using the Wolman pebble count method (Clapcott et al. 2011). The particle size classes used were: clay silt (<0.063mm), sand (0.063-2 mm), small gravel (2-16 mm), medium gravel (16-32mm), large gravel (32-64mm), small cobble (64-128 mm), large cobble (128-256 mm), boulder (>256 mm) and bedrock.

The amount of fine resuspendible sediment in the streambed was qualitatively assessed using the Shuffle Index (method 5 in Clapcott et al. 2011). This was applied in runs where the flow was between 0.2 and 0.6 m/sec and depth was between 20 and 50 cm. The index measures the extent to which a tile placed on the stream bed is obscured after vigorous disturbance: Score 1 = 'No or small plume; Score 2 = 'Plume briefly reduces visibility at tile'; Score 3 = 'Plume partially obscures tile but quickly clears'; Score 4 = 'Plume partially to fully obscures tile but slowly clears'; Score 5 = 'Plume fully obscures tile and persists even after shuffling ceases'.

A photograph was taken of each sample site (see Appendix 1).

2.4 Macrophytes

Aquatic macrophyte cover was assessed over a 30m reach at each site using the rapid assessment protocol in the 'Regional Guidelines for Ecological Assessment of Freshwater Environments: aquatic plant cover in wadeable streams' (Collier et al. 2014). This involved assessing emergent and submerged macrophyte cover and type occupying a one metre wide belt across the stream at five transects spaced along the reach being assessed.

The following metrics were calculated:

- Macrophyte Total Cover (MTC): This reflects the extent and cover over the bottom. It is calculated as: $\{\Sigma$ (% emergent + % submerged)} / 5 transects.
- Macrophyte Channel Clogginess (MCC): This reflects the extent and cover through the water column. It is calculated as: $\{5$ (%emergent + %surface reaching) + (% below surface * 0.5) $\}$ / 5 transects.

2.5 Periphyton

2.5.1 Periphyton cover (including benthic cyanobacteria)

Periphyton cover was assessed at each of the five transects at each site. Periphyton cover was visually assessed at each site using the method and field sheets in Collier et al. (2014) which is based on the Rapid Assessment Method 2 (RAM2) in Biggs and Kilroy (2000). It involves assessing periphyton cover and type on a total of 25 points on five transects located at least 5m apart. For each transect five rocks were assessed across the stream. Each transect approximately corresponded with the location of replicate macroinvertebrate samples. Benthic cyanobacteria were separately identified as part of the periphyton visual assessment.

The sites were searched for any heterotrophic growth and these were recorded at the same time as the periphyton survey. None were found at any site.

The following ecological indices were calculated from the periphyton cover: 1

- Periphyton Enrichment Index (PEI). This is a measure of nutrient enrichment. Higher scores reflect greater algal cover by periphyton categories indicative of nutrient enrichment. The index is a measure of the relative abundance of different types of algae and thus any transects without periphyton are ignored in the calculation. It is calculated as: PEI = {[∑(% cover in each category per transect * Indicator score) / Total % cover per transect] / No. transects with periphyton} *11.
- Periphyton Proliferation Index (PPI). This is an indicator of biomass. It is the percent of total cover by long filaments and thick mats. It is strongly related to % EPT taxa.
- Periphyton Sliminess Index (PSI). This is an indicator of biomass. It is the weighted percent cover of each thickness category and is calculated as: PSI = {(%Thin mat/fil) + (%Short filaments * 2) + (% Medium mat * 3) + (% Long filaments * 4) + (%Thick mat*5)} / 5. It is strongly related to macroinvertebrate diversity and condition.
- Periphyton Weighted Composite Cover (Peri WCC): This is a measure of periphyton cover of mats >3mm thick and filaments green algae > 2cm long. PeriWCC = % filamentous algae + (% mat algae/2) (Matheson et al 2012).

2.5.2 Periphyton biomass and identification

Periphyton biomass was assessed using the Quantitative Method 1b (QM-1b) in Biggs and Kilroy (2000). Five replicate periphyton samples were collected from large gravel / cobbles at each site (one rock per transect). This involved removing all periphyton from a set area of 6cm diameter on the surface of gravels / cobbles representative of periphyton cover on stable substrate at each of five transects.

On 10 October five (5) replicates were collected from each site on Otauira Stream, Donald Creek 100m u/s and Donald Creek 650m d/s (area sampled of 0.002827 m² per replicate and 0.0141372 m² per site). For Donald Creek sites 25m upstream and 60m downstream a single replicate was collected which consisted of periphyton scraped from three and four cobbles respectively (sample area of 0.008482 m² and 0.01131 m^2 respectively).

On 1 November a single replicate was collected from each site, each periphyton biomass replicate consisted of 5 cobbles with 6cm diameter area scraped (total area sampled of 0.0141372 m² per site).

Periphyton biomass samples were frozen and sent to the laboratory for analysis. Samples collected on 10 and 11 October were analysed for chlorophyll *a,* samples collected on 1 November were analysed for chlorophyll *a* and Ash Free Dry Mass (AFDM). Chlorophyll *a* was extracted using the hot ethanol method as described in Biggs and Kilroy (2000).

The results were compared with the NZ Periphyton Guidelines (Biggs 2000). The ratio of AFDM to chlorophyll *a* was used to calculate the autotrophic index (a measure of organic enrichment).

A single bulked sample was collected from each site and analysed for periphyton community composition. The bulked sample consisted of a sub-sample from each rock sampled for periphyton biomass. Samples were refrigerated and sent to the laboratory for identification. The relative

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 1 Note that Collier et al. (2014) modified the formula for calculating PEI from that in the previous version (Collier et al. 2007). Furthermore, the periphyton scoring differs from the scores in Biggs & Kilroy (2000) because they have been subtracted from 10.

abundance was assessed using the method in Biggs and Kilroy (2000) i.e. using a relative scoring system from 1 (rare) to 8 (absolutely dominant/monoculture).

The NZ Provisional Periphyton Guidelines (Biggs 2000) sets guidelines to maintain 'trout habitat and angling' as peak biomass <35 g AFDM/m² (Ash Free Dry Mass), or <200 mg chlorophyll a/m^2 for diatoms/cyanobacteria dominated communities and <120 mg chlorophyll $a/$ m² for filamentous dominated communities. Periphyton cover guidelines are set at <30% filamentous algae over the stream bed. Guidelines set for maintaining benthic biodiversity are < 50 mg chlorophyll a/m^2 .

Matheson et al. (2012) proposed an alternative index of cover called the Periphyton Weighted Composite Cover (Peri WCC). They proposed an aesthetic nuisance guideline of >30% Peri WCC, but also identified that a Peri WWC of <40% corresponded to macroinvertebrate communities in 'good' condition (i.e. QMCI of >5 and MCI >100).

2.6 Macroinvertebrates

The use of macroinvertebrates for assessing the condition of streams is widespread in New Zealand and overseas. The structure and composition of macroinvertebrate communities is a good indicator of stream condition as they are found in almost all freshwater environments, are relatively easy to sample and identify, and different taxa show varying degrees of sensitivity to pollution.

Aquatic macroinvertebrates were collected from riffle/run habitat using a Surber sampler and following Protocol C3 of Stark et al. (2001). All macroinvertebrate samples were preserved in alcohol and processed using Protocol P3 (full count with sub-sampling option) of the Protocols for sampling macroinvertebrates in wadeable streams (Stark et al. 2001).

On 10 October, five (5) replicates were collected from each site on Otauira Stream, Donald Creek 100m u/s, Donald Creek 60m downstream and Donald Creek 650m d/s with each replicate consisting of at least two Surber samples (area sampled > 1 m² per site)². For the Donald Creek site 25m upstream a single replicate was collected which consisted of six Surber samples bulked into a single container (total sample area of 0.6 m^2 for the site).

On 1 November, a single replicate was collected from each site, this replicate consisted of five (5) Surber samples collected from along the reach (a total sample area of 0.5m²).

Macroinvertebrate results were expressed on an areal basis i.e. per square metre. The following ecological indices were calculated to assess the biological health of the river and potential effects on the stream ecology:

- Taxa Richness: This is a measure of the types of invertebrate taxa present in each sample.
- EPT richness and EPT abundance (Ephemeroptera-Plecoptera-Trichoptera). This measures the number of pollution sensitive mayfly, stonefly and caddisfly (EPT) taxa in a sample excluding *Oxyethira* and *Paroxyethira*.
- % EPT abundance.
- Macroinvertebrate Community Index (MCI). The MCI is an index for assessing the water quality and 'health' of a stream using the presence/absence of macroinvertebrates (Stark 1985).

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 2 See results for actual area sampled at each site.

 Quantitative MCI (QMCI). The QMCI is similar to the MCI but is based on the relative abundance of taxa within a community (Stark 1993, Stark 1998).

The MCI and QMCI reflect the sensitivity of the macroinvertebrate community to pollution and habitat change, with higher scores indicating higher water quality. Generally accepted water quality classes for different MCI and QMCI scores and soft-bottomed version are shown in Table 2.2.

Table 2.2: Suggested quality thresholds for interpretation of the MCI & QMCI from Stark (1998)

In addition to sampling benthic macroinvertebrates, the aquatic vegetation and soft sediment within the sample reaches were searched for kākahi (freshwater mussel) and the presence of bivalves such as fingernail clams (*Sphaerium* sp). A kick-net was used to search soft sediment for the fingernail clam. Bivalves are known to be particularly sensitive to the presence of ammonia, so their presence in a stream has implications for waste water discharges.

2.7 Fish survey

Fish were recovered by using baited fyke nets and baited gee minnow traps placed in the stream overnight on 10 to 11 October 2016. Six fine-mesh fyke nets and 6 Gee-minnow traps were placed in each stream on consecutive nights. The survey reach in Donald Creek spanned 140m and the survey reach in Otauira Stream spanned 200m.

The fyke nets were fine mesh (mesh size ca. 4mm) with net dimensions of: 6 hoops, with 60cm wide front D mouth, and 3m long trap and 5m long leader (Joy et al. 2013). The gee-minnow traps had a 4mm mesh size and 2.5cm diameter mouth. All nets and traps were baited with cheese.

2.8 Water Quality

A single water quality grab sample was collected from each site at the time of the survey and a field meter was used to measure water temperature, electrical conductivity and dissolved oxygen. Water samples were sent to Hill Laboratory and analysed for the following variables: pH, total suspended solids (TSS), nitrate-N, nitrite-N, ammoniacal-N, total kjeldahl nitrogen (TKN), dissolved reactive phosphorus (DRP), total phosphorus (TP), cBOD5, and *E.coli* bacteria.

Grab samples were collected prior to walking in the stream for biological sampling. Samples were collected from mid-stream using standard sampling protocols. In addition to this water quality sampling, South Wairarapa District Council collects regular water quality samples from the discharge and Donald Creek.

2.9 Mixing zone survey

The mixing zone at the time of the survey was assessed on 10 October using specific electrical conductivity (EC) as a conservative tracer. EC at 1m intervals across the stream was measured along transects progressively down the stream.

Specific electrical conductivity (EC) is often used as a conservative tracer in mixing studies. The WWTP discharge has higher EC compared to Donald Creek (i.e. about 187.5 µS/cm compared to 136 µS/cm upstream), which allows EC to be used to estimate dilution at points downstream. Dilution was calculated using a mass balance approach over a 'control volume' of the river after mixing. The formula for the dilution factor (D) for downstream sampling points is:

D = (ECeff – ECu/s)/(ECd/s – ECu/s)

Where: $D =$ dilution factor; ECeff = electrical conductivity of the effluent; ECu/s = EC of the river upstream; ECd/s = EC of the river downstream.

2.10 Statistical Analysis

The statistical significance of macroinvertebrate and periphyton indices was compared using an equivalence test in the software 'TimeTrends'. A difference was considered statistically significant if the *p*-value was < 0.05; and a difference was considered "*practically important*" if the difference was +/- 20% change compared to the upstream control sites. Allowing up to a 20% change recognises that habitats can seldom be perfectly matched and even small changes in substrate size, flow and location can impact on macroinvertebrate composition to some extent.

Equivalence tests incorporate both testing of means (using a student t-test) and testing of a meaningful change (interval testing). One advantage of equivalence tests is that increasing the sampling effort may make it either more or less likely that an equivalence hypothesis will be rejected, unlike the statistical test where more data means that the hypothesis is more likely to be rejected.

3 Results

3.1 Habitat and Macrophytes

3.1.1 Donald Creek

Donald Creek has similar water depth (in runs) between the upstream and downstream sites but there was a general increase in stream width (wetted width), and a decrease in water velocity (mid-stream in runs). The size of the substrate also reduced; with large gravel comprising 28% of the substrate at the 100m upstream site and being absent from the 650m downstream site (Table 3.1 and Table 3.2).

The substrate in Donald Creek was loosely packed and easily moved. Sand was moving on the stream bed in Donald Creek sites upstream and 60m downstream of the discharge. Very little substrate was embedded by fine sediment except at the 650m downstream site. Aquatic moss (*Drepanocladus* sp.) was present on some of the cobbles at the 70m downstream site (Table 3.4).

The 650m downstream sites had more resuspendable fine sediment on the stream bed compared to the other sites. This was apparent in the Shuffle Index; the 650m downstream site disturbing the stream bed released fine sediment that obscured the stream bed and took several seconds to clear (Table 3.1). At this site, cattle were in the stream and the stream margins were heavily pugged.

The habitat score was higher for sites within the bush remnant either side of the discharge (25m upstream and 60m downstream). This was because the riparian vegetation provided better diversity of fish cover (e.g. woody debris, root mats, undercut banks, overhanging vegetation), less bank erosion and more riparian

shade. Logs within the stream helped provide hydraulic heterogeneity – particularly at the 60m downstream site (Table 3.1). The immediate riparian margin was dominated by *Apium* sp, pasture grasses, buttercup (*Ranunculus* sp.). In the bush remnant black berry and the pest plant wandering jew (*Tradescantia* sp.) was common.

Donald Creek had low macrophyte cover in the shaded sites (25m upstream and 60m downstream), and moderate macrophyte cover in the unshaded sites (100m upstream and 650m downstream) (Table 3.3). The predominant macrophyte in the stream was water celery (*Apium nodifolrum*) – this is a sprawling emergent macrophyte that was also very common on the riparian margins. The site at 650m downstream had *Apium* sp. present but also had a range of submerged species including: *Elodea canadensis*, *Callitriche stagnalis* (starwort), *Potamogeton crispus* (curled pondweed), some *Glyceria fluitans* (floating sweet grass) and the native *Nitella hookeri*.

Periphyton was more common in the unshaded sites; although the shaded sites did have sufficient light for periphyton to grow (e.g. the 60m ds site had open sky beyond a 30 $^{\circ}$ angle). The periphyton was predominantly thin films or diatom mats (sludge) but tuffs of the red algae *Batrachospermum* sp. was common on the stream bed at 650m d/s site on both sample occasions (Figure 3.1).

3.1.2 Otauira Stream

Otauira Stream is a gravel bed stream. Upstream of the confluence with Donald Creek the runs are wide (10.6m) and shallow (15cm), but downstream it becomes more confined, narrower (7.6m) and deeper (45cm) (Table 3.1).

The substrate on Otauira Stream was dominated by large gravel and small cobbles. Cobbles were more common at the downstream site – perhaps reflecting the confined channel (Table 3.2). Aquatic mosses was common on cobbles at downstream site but were rare upstream – perhaps indicating more stable substrate. At both Otauira Stream sites the substrate in was moderately to tightly packed with very little being embedded by fine sediment.

Both sites on Otauira Stream had reasonably high habitat scores due to very low fine sediments, high diversity and abundance of habitat for invertebrate colonisation (e.g. cobbles, gravel, sand, wood, macrophytes, and leaves), very little bank erosion and moderate riparian width. However habitat scores were lowered by both sites having very little riparian shade.

Otauira Stream had very little macrophyte cover upstream of the confluence with Donald Creek, while the downstream site had about 10% cover of the sprawling emergent *Apium nodifolrum*. *Apium* sp. extended up on to the riparian margins where buttercup (*Ranunculus* sp.), black berry and willow were also common (Table 3.3).

Both sites had relatively little periphyton cover – mostly consisting of thin films or diatom mats. Interesting the upstream site had more periphyton present during the 11 October survey compared to the 1 November survey, while the downstream site had more periphyton present on 1 November.

On 12 October 2016, Otauira Creek had about 1.5 times the flow of Donald Creek³, but by 1 November 2016 the flow in Otauira Stream was considerably less than the flow in Donald Creek. This reflects the numerous springs that contribute to the flow in Donald Creek.

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 3 Based on EC measurements taken at sites upstream and downstream of the confluence after full mixing.

Table 3.1: Summary of habitat variables at sites on Donald Creek and Otauira Stream, 10 – 11 October 2016

Table 3.2: Substrate particle size distribution and Shuffle Index of resuspendable sediment at sites on Donald Creek and Otauira Stream, 10 – 11 October 2016

Si = silt, S =sand, SG=small gravel, MG=medium gravel, LG=large gravel, SC=small cobble, LG=large cobble, B=boulder

Table 3.3: Macrophyte cover at each site on 11 October 2016. Mean of five transects.

3.2 Periphyton and benthic cyanobacteria

3.2.1 Periphyton

There was very little periphyton at any site in Donald Creek during the October 16 survey. Stones had a thin film of algae making them slippery, but there were little visible growths or 'sludges'.

A statistical analysis of the periphyton cover and biomass in Donald Creek during the October survey found no statistically significant changes in PEI, PSI and chlorophyll *a* (due to the small number of replicates relative to the variability within the sites). However the increase in the PPI at the 650m downstream site was statistically significant (t-test *p*-value =0.03) because this was the only site where long filamentous algae was recorded – albeit in small amounts. These were tuffs of the red algae *Batrachospermum* sp. about 3cm long (Figure 3.1).

For Otauira Stream moderate cover of the diatoms *Frustulia* and *Cymbella* sp. were present as 'sludges' on the cobbles at the site upstream of the confluence, but there was very low cover downstream. A statistical analysis of the periphyton cover and biomass from the October survey found that the decrease in PSI and PEI at the downstream site was statistically significant (t-test *p*-value =0.002, strong evidence of a difference >10%). However the lower mean chlorophyll *a* values were not statistically significant due to the small number of replicates relative to the variability (Table 3.4, Table 3.6).

By the time of the November survey Donald Creek had considerably more periphyton cover, but all sites were still below the NZ Periphyton Guidelines for maintaining 'trout habitat and angling', and all sites were well below the PeriWCC guideline of 30% for maintaining aesthetic values . The 25m upstream site had the lowest periphyton cover and biomass. The 650m downstream site had higher periphyton cover as reflected by the indexes PPI and PeriWCC; this was due to patches of the red algae *Batrachospermum* sp. The other cover indexes of PEI and PSI showed no clear difference between upstream and downstream sites (Table 3.5).

The periphyton biomass indicator of AFDM showed a pattern similar to the PEI, with no clear difference between the Donald Creek downstream sites and the 100m upstream sites. However, chlorophyll *a* was noticeably higher at the two downstream sites (Table 3.6). The biomass samples from the 650m downstream sites included a clump of *Batrachospermum* sp. It appeared that the periphyton from the 60m downstream sites was particularly rich in chlorophyll *a*.

During the November survey Otauira Stream had more periphyton cover (EPI and PSI) and higher biomass (AFDM and chlorophyll *a*) downstream of the confluence compared to upstream. This was a reversal of the pattern observed in early October. The change may be due to the interaction between periphyton growth rate and density of macroinvertebrates that graze the periphyton.

Most of the algae species in the periphyton community are typical of mesotrophic to eutrophic conditions (i.e. the diatoms *Nitzschia, Melosira, Cymbella*). *Cymbella* sp. rapidly colonises substrate after floods, which may explain its dominance in the samples from early October. *Achnanthidium* sp. is widespread throughout New Zealand but is often in streams with moderate to high water velocity – perhaps explaining its presence at the 25m upstream site. *Batrachospermum* sp. was common at the 650m downstream site; it is typical of clean cool streams, often spring fed or shady forest streams, and is often associated with high soluble organic carbon such as stable Westland stream (Biggs 2000, Kilroy and Biggs 2000) (Table 3.6).

At sites with smaller substrate (e.g. Donald Creek 650m downstream), the periphyton biomass measurements may have been biased upward compared to the cover measurement because larger rocks were sought for scrapping. These were also the more stable rocks and more likely to have higher periphyton biomass. Green filamentous algae was rare in both stream but was observed on submerged logs at the Donald Creek 60m downstream site.

3.2.2 Benthic cyanobacteria

During the October site visit, occasional patches of benthic cyanobacteria were only found in Donald Creek at the site 650m ds, but in such low cover that it was not found on the periphyton transects. During the 1

November site visit small patches of cyanobacteria were found in Donald Creek at sites 100m upstream of the discharge and 650m downstream.

In Otauira Stream upstream of the confluence, small patches of the benthic cyanobacteria were present on the stream bed on 11 October but cover was so low that it was not detected in the periphyton transects. No cyanobacteria were found on 1 November 2016. Benthic cyanobacteria were not found in Otauira Stream downstream of the confluence on either site visit.

3.2.3 Heterotrophic growths

Each site was searched for the presence of heterotrophic growths and none were found at any site on either sampling occasion.

The Autotrophic Index (AI) is sometimes used as an indicator of organic enrichment. Dissolved organic waste tends to favour the growth of heterotrophic periphyton taxa such as the filamentous bacterium. A shift towards these organisms can be indicated by an increase in the autotrophic index. A higher AI indicates more organic pollution and an AI value >400 can indicate pollution (Collins and Weber 1978 in Biggs 2000). However, care is needed in interpreting AI results if AFDM concentrations are low or if communities are dominated by mucilaginous diatom (e.g. *Gomphoneis, Cymbella, Synedra*) that can have naturally high AI values.

The AI values calculated from samples collected on 1 November give no indication of heterotrophic growths within the periphyton community at sites downstream of the discharge. The only site with a high AI score was on Donald Creek upstream of the discharge and this may be because the low periphyton biomass exaggerates any inaccuracies in the chlorophyll *a* and AFDW measurements (Table 3.6).

Table 3.4: Mean periphyton cover in Donald Creek and Otauira Stream on 11 October 2016. Patches of benthic cyanobacteria occurred at some sites but at low levels not detected by the survey method.

Thickness categories: Thin <0.5mm, medium 0.5-3mm, thick >2mm. Short <2cm, Long >2cm.

Table 3.5: Mean periphyton cover in Donald Creek and Otauira Stream on 1 November 2016.

Thickness categories: Thin <0.5mm, medium 0.5-3mm, thick >2mm. Short <2cm, Long >2cm.

Table 3.6: Periphyton biomass and dominant species in Donald Creek and Otauira Stream, see Appendix 3 for full results of relative species composition and October replicates.

1-Nov-16

Autotrophic Index = AFDM in mg/m² : chlorophyll a in mg/m²

Figure 3.1: Tuffs of the red algae *Batrachospermum* sp. on the stream bed at the site Donald Creek 650m ds.

3.3 Benthic macroinvertebrates

3.3.1 Donald Creek

The aquatic macroinvertebrate community in Donald Creek during October showed taxa richness at the combined downstream sites but a small decline in the percentage of EPT taxa and small decline in the MCI (about 8%). There was no significant difference in the QMCI scores between the combined upstream and downstream sites. The highest MCI and QMCI scores were at the site 25m upstream of the discharge and the lowest 650m downstream, in contrast the lowest QMCI score was at the 100m upstream site. The abundance of EPT taxa and the % abundance of EPT taxa was reasonably low at all of the sites (Figure 3.2, Table 3.7, Table 3.8 and Table 3.9).

The November survey showed a similar pattern, with slightly lower MCI scores at the two downstream sites (up to 7% lower), and no consistent upstream to downstream difference in QMCI scores. However the 60m downstream site had considerably lower QMCI scores than all other sites; this site had considerably fewer *Paracalliope* amphipods, fewer *Potamopyrgus* snails, and fewer EPT taxa but similar abundance of pollution sensitive EPT taxa (e.g. mayflys) as had the upstream sites (Figure 3.2, Table 3.7 and 3.10). The difference in species composition at the 60m downstream site may be partially related to differences in the quantity of habitat sampled (e.g. in October, replicate B had considerably higher abundance of these taxa, Appendix 3).

On both sampling occasions the abundance of mayfly (*Deleatidium* sp.) was low at the 650m downstream site (Table 3.9 and 3.10). This may relate to the smaller substrate size, higher amount of fine sediment and cattle disturbance observed at this site.

3.3.2 Otauira Stream

The macroinvertebrate survey of Otauira Stream during October 2016 found that the downstream site had higher total abundance, higher % EPT abundance and higher QMCI scores, but lower MCI scores and % EPT taxa. *Astrosimulium* (sandfly larvae) and *Orthocladiinae* were very abundant at both sites, but the downstream site was characterised by considerably more *Deleatidium* mayfly and *Potamopyrgus* snails, while the upstream site had more *Oligochaeta* worms (Figure 3.2, Table 3.7, Table 3.8).

The November survey found that the MCI score was still a little lower downstream (7% lower) but all other indices were very similar between the two sites. *Astrosimulium* (sandfly larvae) and *Orthocladiinae* remained numerically dominant at both sites, but the total abundance per square metre had increased.

Also the abundance of *Deleatidium* mayfly (and to a lesser extent *Potamopyrgus* snails) had increased at the upstream site (Figure 3.2, Table 3.7 and 3.10).

Mayfly and snails graze periphyton and the increased abundance of these taxa at the upstream site is likely to explain the reduction in periphyton cover and abundance at the upstream site between October and November surveys (Table 3.7).

3.3.3 Kākahi and fingernail clams

The sample sites on Donald Creek and Otauira Stream were searched for freshwater mussel (kākahi)⁴ and the fingernail clam (*Sphaerium* sp., Sphaeriidae). These species tend to live in sandy or silt areas in streams seeps and lakes. Young kākahi are sometimes found on or amongst macrophytes and many *Sphaerium* sp can climb onto macrophytes to live. Both species are filter feeders and both are sensitive to ammonia concentrations in the water. Kakahi can live for over 30 years, while the fingernail clam (*Sphaerium* sp.) typically lives for 1 to 3 years and can be mature by 3 months.

No kākahi were found at any of the sites. However, the fingernail clam was found at all the Donald Creek sites and in Otauira Stream downstream of the confluence during October. On both sample dates it was most abundant at the 650m downstream site. They were associated with sandy substrate and their distribution within sites was patchy (see Appendix 3).

The downstream sites had lower abundance of *Sphaerium* sp*.* during the November survey compared to October. It is possible that this related to higher total ammonia during November (see below), but it probably reflects the patchy distribution of fingernail clam (preferring coarse sandy substrate) and the smaller area of streambed sampled during the November survey.

 \overline{a} 4 Two species of kākahi are found near Lake Wairarapa, *Echyridella menziesii* and *E. aucklandica.*

Figure 3.2: Median MCI and QMCI scores for Donald Creek and Otauira Stream during the surveys in October and November 2016. The error bars show one standard deviation of replicate samples.

Table 3.7: Summary macroinvertebrate and algae metrics for sampling on 11 October 2016 (median) and 1 November 2016.

EPT exclude *Oxyethira* and *Paroxythira.*

Table 3.8: Results of statistical analysis (equivalence test) between upstream and downstream sites on Donald Creek and Otauira Creek, 11 October 2016.

Donald Creek	Direction $ p$ -value		Evidence strength
Chlorophyll-a		ns	Weak, inconclusive
PSI		ns	Weak, inconclusive
Number of taxa	ds >us	0.02	Moderate but close to the limits
Number of EPT taxa		ns	Weak, inconclusive
% EPT taxa		ns	Weak, inconclusive
% EPT abundance	$ds<$ us	0.04	Moderate but close to the limits
MCI score	$ds<$ us	0.008	Moderate but a trivial difference
QMCI score		ns	No evidence of difference

Table 3.9: Macroinvertebrates per m² in the Otauira Creek and Donald Creek, 11 October 2016. Average of replicates (see Appendix for full results).

3.4 Fish

The fish community in Otauira Stream and Donald Creek was dominated by longfin eel and common bully. A greater diversity of fish was caught in Otauira Stream – longfin, shortfin, common bully, inanga and a small rainbow trout (Table 3.11); this may reflect its closer proximity to Lake Wairarapa. Most of the eel caught in the fyke nets were large, in Otauira Stream seven of the longfin eel were over 1m long, and in Donald Creek four of the longfin eel caught were greater than one metre. The average longfin eel size in both streams (excluding elva) was 84cm, and the largest eel caught was 145cm long (Figure 3.3).

Two medium sized rainbow trout were observed in Donald Creek at the site 650m downstream during the 1 November site visit.

The placement of nets in relation to the WWTP discharge or the Donald Creek confluence didn't appear to make any difference in the number of fish caught.

The wastewater treatment plant operators have recorded a large number of eel living in the wastewater treatment ponds. On one occasion 80 eel were counted in the ponds when an incident resulting in low dissolved oxygen in the ponds, forcing the eel towards the water surface.

Longfin eel and inanga have a threat classification of 'At-Risk – Declining', and their presence in the streams is encouraging (Goodman et al. 2014). The abundance of longfin eel probably reflects the habitat. Large longfin eels prefer deep, slow flowing water. In larger rivers they move from deep water during the day into shallower water at night. More importantly their abundance is related to the presence of in-stream cover (e.g. woody debris), bank cover and low fishing pressure (Garnooth and Brooker 2009).

Table 3.11: Fish caught in Otauira Stream and Donald Creek in October 2016. Each had one night of trapping with six fine-mesh fyke nets and six gee-minnow traps.

Otauira Stream

Donald Creek

eel sizes: 0+ <10cm, small 10-30cm, Medium 30-50cm, Large >50cm common bully sizes: 0+ <2cm, small 2-6cm, Medium 6-15cm, Large >15cm inanga: 0+ <4, small 4-6cm, medium 6-8cm, large >8 cm

Figure 3.3: Two longfin eel in a fyke net set in Donald Creek, October 2016.

3.5 Water Quality

3.5.1 Water quality

Dissolved oxygen (DO) fluctuates diurnally; it is typically high during the day due to photosynthesis by plants and algae, and is low at night due to respiration. The minimum daily DO is particularly important for aquatic life. The dissolved oxygen (DO) concentration in the discharge soon after day break on 11 October was 4.3 mg/L (40% saturation). In Donald Creek, five metres downstream on the true right bank (i.e. only partial mixing) the DO was 8.8 mg/L (81% saturation), and upstream of the discharge it was 9.3 mg/L (88%).

Franklin (2014) developed DO criteria for protecting New Zealand fish communities. Imperative targets protect adults and the majority of fish species. The guideline criterion protects salmonids and early life stages of all species, this requires the instantaneous minimum DO of >5 mg/L and the 7-day mean minimum DO of >6 mg/L. The NPS-FW used an instantaneous minimum DO of >4mg/L as a national bottom-line (Davies Colley et al. 2013). All instream DO measurements were well about these guideline values (Table 3.12).

Water quality samples collected on 12 October 2016 found that upstream of the discharge Donald Creek had high concentrations of soluble inorganic nitrogen (SIN) and reasonably low concentrations of soluble phosphorus (DRP). The discharge elevated the SIN by a relatively small amount (20%), but this was mostly in the form of total ammoniacal nitrogen (ammonium). The ammonium was within guideline values on this occasion. DRP was considerably elevated by the discharge (Table 3.13).

At the time of sampling, upstream of the discharge in Donald Creek, periphyton growth was likely to be limited by DRP but not SIN, while downstream of the discharge nutrient availability was unlikely to place any limit on periphyton growth. In Otauira Stream, both DRP and SIN were likely to be limiting periphyton growth upstream of the confluence, while downstream the DIN was in excess of periphyton growth requirements and the DRP was likely to exert only a little control on periphyton growth.⁵

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⁵ Rier and Steven (2006) found nutrient saturation at dissolved inorganic N (DIN) and DRP concentrations of 0.31 mg/L and 0.038 mg/L respectively; while peak growth rates occurred at DIN and DRP concentrations of 0.086 mg/L and 0.016 mg/L respectively. This is consistent with other studies that have found little growth limitation at DRP concentrations above 0.02 to 0.08 mg/L (Hill and Fanta 2008, Bothwell 1989). Kilroy et al. (2012) found that

On 1 November 2016 water quality samples were collected from Donald Creek as part of regular monitoring by South Wairarapa District Council (sites referred to as 'Longwood'). The concentration of total ammonia and DRP at the downstream sites were higher compared to previous months. The DRP concentration (0.123 mg/L) was sufficiently high to not exert any control on periphyton growth. The total ammonia concentration (0.51 mg/L) was borderline for protection of sensitive species like the fingernail clam (No Observed Effect Concentration (NOEC) of 0.54 mg/L, Hickey 2014), but was within ANZECC guidelines for protection for 95% species protection (i.e. 0.9 mg/L), and within the bottom-line set in the NPS-FM (i.e. annual median of 1.3 mg/L and annual maximum of 2.2 mg/L)⁶.

			Temp.	spec EC	DO	
Date	time	site	(OC)	(uS/cm)	(mg/L)	DO%
		10/10/2016 9:40 Donald 100m us	10.8	136.9	11	100
		10/10/2016 12:10 Donald 25m us	13.9	135.8	10.6	103
10/10/2016 13:10 discharge			14.7	187.3	6.8	67
		10/10/2016 14:30 Donald 60m ds	15.1	139.5	10	100
		10/10/2016 16:05 Donald 650m ds	16.5	143.5	8.8	90
		11/10/2016 12:15 Otauira 70m us	11.6	101	11.5	106
		11/10/2016 14:10 Otauira 100m ds	12.9	125	10	97

Table 3.12: Field measurements from Donald Creek and Otauira Stream on 10-12 October 2016. Temp =Temperature, EC = specific electrical conductivity, DO = dissolved oxygen.

Table 3.13: Water quality of Donald Creek and Otauira Stream on 12 October 2016

		Donald	Donald		
Variable	Guideline	25m us	60 _m ds	Otauira us	Otauira ds
pH	$6.5 - 9.0$	7.4	7.3	7.3	7.2
TSS (mg/L)		\leq 3	4	$<$ 3	3
TN (mg/L)		1.3	1.8	0.16	1.29
$NH4-N$ (mg/L)	0.27/0.46	< 0.010	0.195	0.01	0.048
NNN (mg/L)		1.1	1.12	0.069	0.97
SIN (mg/L)	0.1	1.105	1.315	0.079	1.018
DRP (mg/L)	0.011	0.007	0.055	0.005	0.028
TP(mg/L)		0.01	0.081	0.008	0.041
dissolved BOD5 (mg/L)	2	< 2	< 2	< 2	< 2
E. coli bacteria (cfu/100mL)	260 / 550	579	461	435	649

TSS = total suspended solids, NH_4-N = total ammoniacal nitrogen, NNN = nitrate nitrite nitrogen, TN = total nitrogen, DRP = dissolved reactive phosphorus, TP = total phosphorus, $cBOD₅ = carbonacious biological oxygen demand.$

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periphyton biomass in rivers from the Manawatu region did not exceed guideline values if mean soluble inorganic nitrogen was < 0.1 mg/L.

 6 Guideline values assume pH 8 and 20 $^{\circ}$ C.

Table 3.14: Water quality of Donald Creek 1 November 2016. These sites were sampled by South Wairarapa District Council.

3.5.2 Foam and colour

Small amount of foam were observed in Donald Creek at sites both upstream and downstream of the WWTP discharge, and a small amount of foam was observed in Otauira Stream downstream of the Donald Creek confluence on 11 October (Figure 3.4). Foam was not obvious in the WWTP discharge drain at the time of the survey and it is likely the foam observed was due to the decomposition of aquatic macrophytes (e.g. *Apium sp.*) releasing organic material that foams in the turbulent water.

There was no conspicuous difference in either water clarity or colour between upstream and downstream of the discharge on either sampling occasion. It was difficult to tell the difference between water samples in a clear bottle (see Appendix 1).

Figure 3.4: Foam observed in Donald Creek upstream of the WWTP discharge (top) and in Abbot Creek downstream of the Donald Creek confluence (bottom), 11 October 2016.

3.6 Mixing zone

The mixing zone was assessed by measuring specific electrical conductivity (EC) across the stream width at distances progressively further downstream of the discharge. The discharge enters from the true right bank and was fully mixed across Donald Creek by 45m downstream – downstream of a willow log lying across the stream. Considerable dilution had occurred even within 12m of the discharge; after full mixing the dilution factor was about 11.4, while after only 12m the dilution factor on the true right bank was already 9.5 times (Figure 3.5). The dilution factor at the time of the survey was similar to the annual average of 12 times.

At the time of the survey on 10 October, Donald Creek was flowing at 356 L/s and the discharge was flowing at 34.1 L/s. The EC of the discharge was 187.3 µS/cm. The EC of the Donald Creek immediately upstream was about 135 µS/cm.

The mixing zone is likely to be different during summer when there is considerably less flow in Donald Creek (e.g. the median flow in March is 80 L/s).

The ca. 50m required to obtain full mixing is considered reasonable for this stream because the effluent plume along the true right bank did not interfere with the confluence of any tributary, the habitat value on the true right bank were similar to those on the true left, and the site is difficult for the public to access so the zone of non-compliance is not interfering with recreation or aesthetic values.

Figure 3.5: Specific electrical conductivity at 1m intervals across the stream and from 2m to 45m downstream of the discharge.

4 Discussion

4.1 Summary of the spring survey

4.1.1 Donald Creek

Donald Creek had macroinvertebrate communities indicative of 'fair' water quality and 'probably moderate pollution' (based on MCI scores). Good riparian cover and diversity of hydraulic regimes supported populations of large longfin eel and common bully. Sampling during spring found that sites downstream of the discharge had more periphyton and lower MCI scores, but the magnitude of these effects were relatively small; periphyton biomass was within guideline values and the change in MCI score was <7%. In addition to the discharge, the lower MCI scores at the 650m downstream site would also be partially caused by a smaller substrate size, localised sediment input and direct disturbance by cattle in the stream.

4.1.2 Otauira Stream

Otauira Stream upstream of the confluence with Donald Creek had macroinvertebrate communities indicative of 'fair' water quality and 'probably moderate pollution' based on MCI scores; but QMCI scores indicated poorer conditions. MCI scores were lower downstream of the confluence but other indices had an inconsistent pattern, and in October the QMCI was significantly higher at the downstream site. The fish community was dominated by longfin eel and common bully.

There was an interesting pattern of periphyton cover and biomass being higher upstream of the confluence in October when the abundance of invertebrate grazers (e.g. snails and mayfly) was low, but in November higher densities of invertebrate grazers (>300/m²) corresponded to very low periphyton cover and biomass.

4.2 Comparison with summer surveys

Previous surveys of Donald Creek have been conducted during summer low flow conditions and have found the discharge having considerably more impact on the stream than what was found during our spring

surveys (October and November 2016). Previous surveys undertaken in late summer of 2010 and 2013 found that the discharge caused a significant reduction of all macroinvertebrate metrics. Periphyton cover was also elevated at the downstream sites but not as much as expected – possibly due to shading and the deposition of planktonic algae from the oxidation ponds as a scum on substrate⁷. During these summer surveys the effluent caused a noticeable plume and a small amount of heterotrophic growth (5% cover) was present on the streambed (Coffey 2010, 2013).

The spring surveys found the effect of the discharge to be relatively mild compared to those observed during late summer. There was some increase in periphyton cover and biomass but these were within guideline values to maintain aesthetic values and 'trout habitat and angling' (Biggs 2000). The macroinvertebrate community had slightly lower MCI scores at the two downstream sites (up to 7% lower), and no consistent upstream to downstream difference in QMCI scores (Figure 4.1).

In contrast to late summer, spring surveys found no conspicuous change in water colour or clarity from the discharge, and there were no visible heterotrophic growths present in the stream.

This stark seasonal difference in the effect of the discharge is likely to reflect seasonal differences in stream flow, effluent quality and water temperature. The flow in Donald Creek is highly seasonal with a distinct low flow period from about December to April (inclusive) (Figure 4.2). The flow in Donald Creek at the time of the summer survey on 13 April 2010 and 4 March 2013 was 98 L/s and 50 L/s respectively – providing considerably less dilution than during winter and spring. The higher flow also causes movement of sand on the stream bed which will contribute to scouring of periphyton. Some aspects of effluent quality are also worse during summer, with algae proliferation within the ponds affecting the colour and turbidity of the discharge as found by Forbes (2013). Seasonal algae proliferation within the ponds will also cause more extreme dissolved oxygen fluctuations within the effluent. Furthermore, warmer water during summer months can accentuate stress on stream biota, particularly with relation to impacts from ammonia or low dissolved oxygen (Davies-Colley et al. 2013).

When considering options to reduce the effects of the discharge on Donald Creek, the priority should be to reduce discharges during summer and periods of low flow.

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 7 Some willows were killed in 2013 which appeared to reduce the effect of shading by 2016. Unlike the previous surveys, the 2016 survey found some emergent macrophyte cover at all sites, indicating less difference between sites due to shading.

Figure 4.1: Median MCI and QMCI scores for Donald Creek for surveys in April 2010, March 201, October 2016 and November 2016 (Coffey 2010, 2013). The error bars show one standard deviation of replicate samples. The red horizontal line indicates scores indicative of 'poor' water quality /habitat.

Figure 4.2: Seasonal variation in flow in Donald Creek (2005-2016). The graph shows the median, 50%ile (within the box), 95 %ile error bars and extreme values. The annual median flow is 241 L/s (Butcher 2016). The median flow for the months January to March is about 75 L/s.

5 Conclusions

Sampling during October and November 2016 found that the WWTP discharge was having only a small impact on the periphyton cover and macroinvertebrate communities in Donald Creek. This is in stark contrast to the effects reported during late summer (2010 and 2013). The difference in the effect of the discharge in spring compared to summer is likely to reflect seasonal differences in stream flow, effluent quality and water temperature.

It is recommended that upgrade and land treatment options give priority to reducing the effects of the discharge during summer and periods of low flow.

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Appendix 1: Site photographs

Photo A1: Donald Creek 100m upstream (facing upstream), 11 October 2016.

Photo A2: Donald Creek 25m upstream (facing upstream), 11 October 2016.

Photo A3: Donald Creek 60m downstream (facing upstream), 11 October 2016.

Photo A4: Donald Creek 650m downstream (facing upstream), 11 October 2016.

Photo A5: Donald Creek 650m downstream (facing upstream), 1 November 2016. Cattle pugging on stream margins.

Photo A6: Water clarity in Donald Creek on 11 October 2016, from left to right, Donald Creek upstream, Donald Creek downstream, and the effluent before mixing.

Photo A7: Substrate in Donald Creek 100m upstream, algae on large gravel (1 November 2016)

Photo A8: Small substrate in Donald Creek 650m downstream, with patches of the red algae *Batrachospermum* sp. (1 November 2016).

Photo A9: Otauira Stream at the confluence with Donald Creek (left hand side), 11 October 2016.

Appendix 5: Periphyton cover

Table A1: Periphyton cover in Donald Creek and Otauira Stream, 1 November 2016.

Donald Creek 100m us

Donald Creek 25m us

Donald Creek 60m ds

Donald Creek 650m ds

Otauira Stream 70m us

Otauira Stream 100m ds

Appendix 3: Periphyton

Table A2: Periphyton community composition. Relative abundance scoring system from 1 (rare) to 8 (absolutely dominant/monoculture) based on Biggs and Kilroy (2000).

1-Nov-16

Table A3: Periphyton chlorophyll a in Donald Creek and Otauira Stream on 11 October 2016

Appendix 3: Macroinvertebrate results

Macroinvertebrates per square metre, 11 October 2016.

